Formation and transport of corrosive water in the Pacific Arctic region

Jessica N. Cross, Jeremy T. Mathis, Robert S. Pickart, Nicholas R. Bates

Ocean acidification (OA), driven by rising anthropogenic carbon dioxide (CO₂), is rapidly advancing in the Pacific Arctic Region (PAR), producing conditions newly corrosive to biologically important carbonate minerals like aragonite. Naturally short linkages across the PAR food web mean that species-specific acidification stress can be rapidly transmitted across multiple trophic levels, resulting in widespread impacts. Therefore, it is critical to understand the formation, transport, and persistence of acidic conditions in the PAR in order to better understand and project potential impacts to this delicately balanced ecosystem. Here, we synthesize data from process studies across the PAR to show the formation of corrosive conditions in colder, denser winter-modified Pacific waters over shallow shelves, resulting from the combination of seasonal terrestrial and marine organic matter respiration with anthropogenic CO₂. When these waters are subsequently transported off the shelf, they acidify the Pacific halocline. We estimate that Barrow Canyon outflow delivers ~2.24 Tg C yr⁻¹ to the Arctic Ocean through corrosive winter water transport. This synthesis also allows the combination of spatial data with temporal data to show the persistence of these conditions in halocline waters. For example, one study in this synthesis indicated that 0.5–1.7 Tg C yr⁻¹ may be returned to the atmosphere via air-sea gas exchange of CO₂ during upwelling events along the Beaufort Sea shelf that bring Pacific corrosive halocline waters to the ocean surface. The loss of CO₂ during these events is more than sufficient to eliminate corrosive conditions in the upwelled Pacific halocline waters. However, corresponding moored and discrete data records indicate that potentially corrosive Pacific waters are present in the Beaufort shelf break jet during 80% of the year, indicating that the persistence of acidic waters in the Pacific halocline far outweighs any seasonal mitigation from upwelling. Across the datasets in this large-scale synthesis, we estimate that the persistent corrosivity of the Pacific halocline is a recent phenomenon that appeared between 1975 and 1985. Over that short time, these potentially corrosive waters originating over the continental shelves have been observed as far as the entrances to Amundsen Gulf and M’Clure Strait in the Canadian Arctic Archipelago. The formation and transport of corrosive waters on the Pacific Arctic shelves may have widespread impact on the Arctic biogeochemical system and food web reaching all the way to the North Atlantic.

1. Introduction

Ocean Acidification (OA) has a significant effect on marine biogeochemistry through the absorption of anthropogenic carbon dioxide (CO₂), both by lowering the pH of the surface ocean and the saturation states (Ω) of biologically important calcium carbonate (CaCO₃) minerals (e.g. Caldeira and Wickett, 2003; Sabine et al., 2004; Bates et al., 2014a). Seawater undersaturated with respect to CaCO₃ (i.e. Ω < 1) is potentially corrosive for these minerals, and the emergence of corrosivity resulting from OA could have potentially negative consequences for calcifying organisms. Physiological impacts are dependent on organismal sensitivity and exposure to low pH and corrosive conditions.

Previous work has identified the Pacific Arctic Region (PAR; Fig. 1) as an OA hotspot (e.g. Fabry et al., 2009). Cooler water temperatures, ocean currents, and unique biogeochemical preconditions the regional carbon system to have a high sensitivity to anthropogenic CO₂ (Bates and Mathis, 2009a, 2009b). Without anthropogenic CO₂, these natural preconditioning processes do not produce seasonal or chronic undersaturations (Mathis et al., 2011a, 2011b). However, given this natural vulnerability, even small amounts of anthropogenic CO₂ can cause large...
chemical changes and result in unnatural corrosivity that defines an OA event (Steinacher et al., 2009). In extremely corrosive hotspots, evidence of carbonate mineral dissolution has already been observed (Cross et al., 2013).

Future projections indicate that the duration, intensity, extent, and frequency of OA events in the PAR are likely to increase as anthropogenic CO$_2$ continues to build up (Steinacher et al., 2009; Mathis et al., 2015b). Rapid onset of persistently acidified conditions can have important consequences in highly sensitive systems like the PAR. The unprecedented pace of global anthropogenic acidification (Hönisch et al., 2012) can overwhelm natural physiological plasticity and population resilience, especially for larger species with lower reproductive rates (Hofmann et al., 2010; Boyd, 2011; Kelly and Hofmann, 2012). Global biogeochemical models have suggested that surface water corrosivity in the Chukchi and Beaufort Seas will pass outside the range of natural variability within the next 10–15 years (Mathis et al., 2015b), representing chronically acidified conditions. Initially, seasonal processes will mitigate these conditions during part of the year, but even these brief periods of respite will be overwhelmed by mid-century (Mathis et al., 2015b) as the system becomes acutely acidified. Given these expected changes, Punt et al. (2014) projected significant declines in populations of commercially important crab species in the Bering Sea, and a recent economic study estimated that current rates of OA already represent a significant risk to economic, cultural, and food securities in Alaskan communities (Mathis et al., 2015a).

As a result of these newly-emergent biogeochemical changes, ecosystem sensitivity, and rapid future progression of acidification, the PAR is commonly referred to as a bellwether for human-induced acidification (e.g. Fabry et al., 2009). Other rapid environmental changes like OA have already caused major shifts in Arctic marine ecosystems: for example, ocean warming, changes in atmospheric and oceanic circulation patterns, and sea-ice losses (Wassmann et al., 2011; Kulkarni et al., 2012; Grebmeier and Maslowski, 2014) have caused populations of temperate taxa to increase (Overland and Stabeno, 2004; Grebmeier et al., 2006) and primary production to rise (Leu et al., 2011; Tremblay et al., 2012; Arrigo and van Dijken, 2015). In particular, these changes are already visible in the diet and body condition of upper trophic marine mammals and birds (Moore and Stabeno, 2015). Because of these close linkages and tight biophysical coupling, many upper trophic species are sentinels for environmental changes (Moore et al., 2014; Moore and Gulland, 2014), and OA impacts could similarly propagate far up the food chain.

While it is clear that the ecosystem in the PAR is likely vulnerable to present and future OA conditions, the research community is just beginning to explore these vulnerabilities in detail. In part, these studies start with an assessment of the duration, intensity, and extent of current ecosystem exposure to corrosive conditions (e.g. Sigler et al., 2017). It is also critical to estimate the role of anthropogenic CO$_2$ in this corrosivity in order to attribute this exposure directly to OA. In some areas, corrosive conditions may be the natural result of regional biogeochemical cycles, and the ecosystem is likely correspondingly adapted (e.g. Hall-Spencer et al., 2008), while in others anthropogenic CO$_2$ pushes the system across key biogeochemical thresholds like the carbonate mineral saturation horizon (e.g. Feely et al., 2009). Understanding these components of OA exposure provides a baseline for future laboratory studies to assess species- and population-specific vulnerabilities (Sigler et al., 2017). Supporting these regional laboratory studies is especially important in the PAR, where many genetically distinct Arctic species and populations have yet to be assessed and may exhibit different responses than their sub-Arctic or tropical counterparts, similar to variability in response to temperature changes (e.g. Hollowed et al., 2013).

To initiate that effort, here we synthesize sub-regional process studies across four major recent programs in the East Siberian, Chukchi, and Beaufort Seas to provide this baseline assessment. We focus first on the unique mechanisms that contribute to natural carbon accumulation and corrosive water formation across the continental shelves of the PAR, exploring new insights into shelf-wide acidification mechanisms. We then combine these results with new moored data records that show the temporal persistence of acidified conditions during and after transport off the continental shelves. Lastly, we estimate the direct contribution of anthropogenic acidification to these corrosive conditions in order to assess the timing and rate of onset of chronic corrosivity. Overall, this synthesis highlights that anthropogenic acidification has drastically increased the duration, intensity, and extent of OA events in the PAR over the last several decades, and that present-day mitigation processes (e.g. upwelling) are relatively minor in comparison to the magnitude of acidification that has already occurred. We conclude by exploring the potential consequences of the rapid rate of onset of acidified conditions for PAR ecosystems over the last several decades, and how chronically acidified conditions may be impacting the present ecosystem structure.

2. Methods

2.1. Discrete data

The dataset used in this study was compiled from four Pacific-Arctic research programs that took place in summer and autumn between 2008 and 2012: the International Siberian Shelf Study (ISSS) in August/September 2008; the Russian-American Long-term Census of the Arctic (RUSALCA) in September 2009; the Impacts of Climate on Ecosystems and Chemistry of the Arctic Pacific Environment (ICESCAPE) in June/July 2010 and 2011; and the Arctic Observing Network (AON) Ocean Acidification Assessment (AON-OA) in October 2011 and 2012 (Fig. 1; Table 1). Together, these measurements span the Pacific inflow from
Table 1

<table>
<thead>
<tr>
<th>Program</th>
<th>Dates</th>
<th>n</th>
</tr>
</thead>
<tbody>
<tr>
<td>ISSS 2008</td>
<td>15 Aug–26 Sept</td>
<td>968</td>
</tr>
<tr>
<td>RUSALCA 2009</td>
<td>4 Sept–29 Sept</td>
<td>1200</td>
</tr>
<tr>
<td>ICESCAPE 2010</td>
<td>18 June–16 July</td>
<td>1167</td>
</tr>
<tr>
<td>ICESCAPE 2011</td>
<td>28 June–24 July</td>
<td>1652</td>
</tr>
<tr>
<td>AON-OA 2011</td>
<td>6 Oct–19 Oct</td>
<td>282</td>
</tr>
</tbody>
</table>

Bering Strait; mixing and modification over the Chukchi, East Siberian, and Beaufort Sea shelves; and transport into the Canada Basin and towards the Canadian Arctic Archipelago.

During each of the programs considered here, discrete samples were collected for dissolved inorganic carbon (DIC) and total alkalinity (TA), alongside additional variables such as temperature, salinity, nutrients, and dissolved oxygen concentration, much of which has been previously published (e.g. ISSS: Anderson et al., 2009; Semiletov et al., 2016; RUSALCA and ICESCAPE: Bates et al., 2013, Bates, 2015; AON-OA: Mathis et al., 2012; Evans et al., 2015). Each program used similar, highly precise sampling and analysis methods that conformed to community best practices (Dickson et al., 2007). During all programs, DIC and TA measurements were routinely calibrated using Certified Reference Material (CRM; supplied by A.G. Dickson, Scripps Institute of Oceanography). Each program reported statistical accuracy ≤ 2 μmol kg⁻¹ (~0.1%). For this synthesis, extended carbon system variables (e.g. pH; partial pressure of carbon dioxide, pCO₂; aragonite saturation state, Ω₉ₐ) were calculated uniformly from the discrete temperature, salinity, phosphate, silicate, DIC and TA observations reported by each program using the CO₂calc software (Robbins et al., 2010). This application relies on specified carbonate dissociation constants. While there are many sets of equilibrium constants commonly used in the literature, we applied those from Millero et al. (2006). These were shown to provide the best comparison between calculated system variables and discrete samples for pCO₂ in the Arctic (Evans et al., 2015). Error in the calculation of Ω₉ₐ was estimated to be ± 0.10, after Dickson (2010) and Hydes et al. (2010).

2.2. Beaufort slope mooring

We also present new results here from the AON-OA moored time series, located at the BS3 mooring (Fig. 1). This time-series site initiated during the Western Arctic Shelf-Basin Interactions program as part of a 7-mooring array spanning the Beaufort Sea shelfbreak and continental slope at 152°W. The original array helped to identify various processes associated with the Beaufort Sea shelfbreak current, including eddy activity (Spall et al., 2008a, 2008b) and upwelling events of Pacific halocline water that have visible impacts on primary production and CO₂ storage (Pickart et al., 2013; Mathis et al., 2012). Subsequent work has shown that the BS3 site alone, located in the center of the jet, can provide a holistic view of the boundary current behavior (Pickart et al., 2009). Accordingly, we assume that the data collected at BS3 are indicative of boundary current conditions in the Beaufort Sea. This in turn can help provide insights into the transport of carbon via winter-modified Pacific Water (Pacific Winter Water, PWW) at the edges of the East Siberian and Chukchi Sea shelves and the interaction with the Pacific halocline layer of the Arctic Ocean. Note that here we discuss the Pacific halocline as a single layer, given that it is difficult to differentiate variability in carbonate corrosivity between the upper and lower Pacific haloclines.

In October of 2012, the AON-OA mooring was deployed in 147 m of water, instrumented with a 300 kHz ADCP, SBE39 MicroCat, and ULS-5 ice profiling sonar at 35 m; and a SAMI pCO₂ sensor and SBE39 MicroCat located at 127 m. Data were collected from 18 October 2012–15 October 2013. Discrete subsurface (> 35 m) DIC and TA data collected during the AON-OA cruise in 2012 were used to derive an empirical relationship between calculated pCO₂ and Ω₉ₐ, such that Ω₉ₐ = 237.12 × pCO₂⁻⁰.⁶⁸⁶ (r² = 0.97; n = 551) (Fig. 2). According to this empirical fit, the data show that when pCO₂ values reached 545 ± 33 μatm (~6% error), Ω₉ₐ crossed the saturation horizon. Error was calculated as one standard deviation from the mean fit. It should be noted that the discrete data used to calculate this relationship were collected during October, and likely represent warmer temperatures than the annual mean. Given that the carbon system is extremely sensitive to temperature, it is important to note here that our empirical relationship is likely conservative—that is to say, during extremely cold periods, a lower pCO₂ value may correspond to the saturation horizon.

This conservative empirical relationship was applied to the pCO₂ time-series from the AON-OA mooring, creating a corresponding time-series for Ω₉ₐ according to the method of Mathis et al. (2013). This time-series shows the seasonal evolution of pCO₂ and Ω₉ₐ as PWW is transported off the shelf, enters the boundary current, and ventilates the Pacific halocline. The physical data from the AON-OA mooring were also used to create a record of wind-driven upwelling events via a graphical user interface (GUI; Supplemental Fig. 1). Upwelling events were defined using the alongcoast wind speed time-series from the Pt. Barrow, AK meteorological station, the along-isobath current velocity, and salinity averaged over the bottom 50 m of the water column. Upwelling is associated with an enhancement of the easterly winds, reversal of the shelfbreak jet (i.e. flow to the west), and an increase of near-bottom salinity (see Pickart et al., 2009 and Schulze and Pickart, 2012 for details). Although not considered here, other quantities such as satellite ice concentration and ice velocity from the mooring can be included in the GUI (e.g. Lin et al., 2016).

3. Results

The circulation and water masses of the PAR have been extensively reviewed in this volume (e.g. Bond et al., 2018; Citta et al., 2018; Logerwell et al., 2018; Moore et al., 2018; and Wang et al., 2018). To put the results of this synthesis study in context, we provide a schematic of the PAR circulation in the lower panel of Fig. 1, similar to those seen.
in Logerwell et al., 2018 and Moore et al., 2018, Pacific Water (PW) preconditioned over the Bering Sea shelf enters the Chukchi Sea through three primary branches; the Alaska Coastal Current (ACC), driven by warmer, fresh water that flows along the Alaskan coast; a middle branch that flows through a topographic depression over the central Chukchi Shelf known as the Central Channel; and the western branch, which flows through Herald Canyon. During summer, most PW throughflow enters the Chukchi through the ACC and the Central Channel current branch (Gong and Pickart, 2015); during winter, transport over Hope Valley and through Herald Canyon comprises nearly 50% of the total throughflow (Woodgate et al., 2005b; Spall, 2007). Water also enters the Chukchi Sea from the East Siberian Sea shelf, which undergoes its own unique preconditioning processes. This water joins the westernmost branch of PW inflow. Over the northern Chukchi Sea shelf, flows are much slower and more complex, particularly near Hanna Shoal (e.g. Pickart et al., 2016).

These slow flows allow for extensive physical and biogeochemical preconditioning (see also Moore et al., 2018). Overall, the seasonal ice cycle partitions this inflow into two distinct water masses. Warming, ice melt and terrestrial freshwater runoff create a buoyant surface water layer that forms the Polar Mixed Layer (PML). Meanwhile, brine rejection and cooling occurring during ice formation produces a dense, saline water mass at the subsurface known Pacific Winter Water (PWW; Pickart et al., 2005). Shelf-basin interaction occurs as outflow through Barrow Canyon, the formation of eddies by instabilities near the shelf-break, and upwelling of deep waters. The variable Chukchi shelf-break jet (dotted line, lower panel Fig. 1) and the Beaufort shelf-break jet then carry this shelf-modified water towards the Canadian Archipelago.

As a ‘first look’ into the corrosivity of the PAR water masses, we present the relationship between temperature, salinity, and ΩAr across the data record for this synthesis in the upper panel of Fig. 3. A vector diagram showing the different water masses of the PAR by temperature and salinity is shown as an inset. In the deep waters of the Canada Basin, CO2 concentrations are naturally very high, resulting in low ΩAr typical of deep waters across the world oceans. The warmer Atlantic halocline layer is supersaturated with respect to ΩAr (warm colors), while ΩAr undersaturations (cool colors) were observed in association with the low-temperature extremes typical of PWW and the Pacific halocline layer of the Arctic Ocean basin. Here, we define the temperature / salinity range of PWW as -1.8 °C < T < -1.4 °C, 32.8 < S < 33.5 (e.g. Woodgate et al., 2005a; Pickart et al., 2016). In this range, ΩAr values averaged 0.97. Note that ΩAr undersaturation was particularly pronounced during October of 2012 (average ΩAr = 0.78; lower panel, Fig. 3, black grid lines). Surface water conditions across the synthesized data record showed ΩAr supersaturations in warmer PML waters. Some ΩAr undersaturations were present at the low-salinity extremes, associated both with river water and ice melt sources. (Note that these waters are typically corrosive endmembers; e.g. Yamamoto-Kawai et al., 2009; Mathis et al., 2011a, 2011b). These endmembers were easily identified in their respective datasets based on location (nearshore / offshore), depth (near surface), salinity, temperature, and TA.

Of all these water masses, transport of corrosive waters through the Pacific halocline layer has the potential to impact the greatest ecological area. Of the various water masses formed from the Pacific inflow, these ΩAr undersaturations are both the most persistent and the most widespread, as shown from the synthesized data above. These waters are regularly upwelled over the Beaufort Sea shelf and transported into the inlets of the Canadian Arctic Archipelago (Fig. 1; Pickart et al., 2013; Mathis et al., 2012; Shadwick et al., 2011). The Pacific halocline also penetrates deep into the Canada Basin, indicating the storage of a high volume of corrosive water that can be maintained for extensive periods of time in this stable water layer. In the following sections, we explore the biogeochemical modification occurring over the shelves of the PAR that create these ΩAr undersaturations and the conditions that favor their persistence, ultimately resulting in ΩAr undersaturation in the Pacific halocline layer.

3.1. Corrosivity in the Pacific inflow

During RUSALCA, supersaturations of ΩAr were observed over the southern Chukchi Shelf just north of Bering Strait. Identifying the mechanism contributing to these CO2 supersaturations is extremely difficult given the lack of knowledge of carbon transport through Bering Strait. The Pacific inflow is known to be extremely nutrient-rich, supporting extensive seasonal phytoplankton photosynthesis (e.g. Codispoti et al., 2005) which could easily draw down CO2 concentrations and favor ΩAr supersaturation. It is possible that mixing also acts to deplete CO2 accumulation outside the production season. The warm, fresh, vertically mixed waters of the ACC are likely supersaturated due to extremely efficient sea-air CO2 exchange, but it would be surprising for this process to extend to the central channel branch of the Bering Strait inflow. Given that very little carbon data are available north of St. Lawrence Island and south of Point Hope, discussion of these mechanisms are merely speculative.

Other work (Mathis et al., 2011a, 2011b; Cross et al., 2012) has hypothesized that intense biogeochemical preconditioning occurring over the southern Bering Sea shelf may contribute seasonally corrosive water generated in fall and winter to the Chukchi Shelf as it flows northwards. ΩAr undersaturations emerge in stratified bottom waters of the Bering Sea as early as June. These corrosive conditions intensify and expand geographically through fall (Mathis et al., 2013). During the Chukchi Sea occupations included in this synthesis, some of this
corrosive Pacific-origin winter-modified water should have been present over the shelf, but mixing and dilution during northward advection may have mitigated these ΩAr undersaturations. It is possible that sea-air CO2 outgassing also contributed to the mitigation of these corrosive conditions, but it should be noted that limited CO2 outgassing has only been observed over the Bering Sea Shelf during vertical overturning typical of the stormy autumn season (Cross et al., 2014).

3.2. Formation of corrosive conditions in the East Siberian Sea: the terrestrial pump

The mechanisms of biogeochemical modification over the East Siberian and Chukchi Sea shelves are better resolved than carbon transport through Bering Strait. Moving northward along the flow path of Pacific waters, the first ΩAr undersaturations encountered in this dataset were found through the RUSALCA program record, in the westernmost branch of the Bering Strait inflow at 68°N and 70°N. In particular, ΩAr undersaturations were observed in bottom waters on the western side of Herald Canyon and around Wrangel Island, which receives some source waters from the East Siberian Sea (Bates, 2015). It is possible that this corrosivity derives both from the respiration of locally produced and terrestrial organic matter. In order to investigate the formation and subsequent transport of this corrosivity, we will turn first to biogeochemical cycling that occurs in the East Siberian Sea.

The ISSS program and subsequent work have substantially added to inorganic carbon system observations in the Russian Arctic over the past decade (Anderson et al., 2010, 2013; Semiletov et al., 2016), especially in relation to the remineralization of terrestrial organic matter. The Lena, Indigirka, and Kolyma Rivers all drain into the Laptev and East Siberian Seas (see Fig. 1), comprising almost a third of the large river discharge in the Arctic Ocean basin. The chemistry of this terrestrial input is also unique: newly eroding soil and permafrost carbon from recent coastal erosion (e.g. Charkin et al., 2011; Semiletov et al., 2016) give the Lena River the highest flow-weighted average contribution of dissolved organic matter of any large Arctic river (762 μM; Cooper et al., 2008).

Much of this organic matter is respired near the coast; a sub-surface respiration signal is apparent near the Lena River outflow, indicated by depleted bottom water oxygen saturation levels and undersaturation with respect to ΩAr (see Fig. 4). In another study using these data, Semiletov et al. (2016) identified a terrestrial origin for this remineralized CO2 using inorganic carbon isotopic data. The data shown here support that analysis; although pCO2 values in the overlying surface plume were low (mean pCO2 158 ± 35 μatm for S < 25) relative to atmospheric values (~386 μatm; Anderson et al., 2010; Tans and Keeling, 2016), oxygen saturation values were approximately 100%, indicating equilibration with the atmosphere. Oxygen saturation values typically exceed 100% during a bloom. The absence of a strong primary production signal confirms that autochthonous marine production is unlikely to be the source of this respired carbon in the bottom layer. Instead, the low surface pCO2 values were likely created by dilution with river water.

In extremely shallow areas, mixing driven by winds, tides, and seasonal storm events can overturn the entire water column (e.g. Coachman, 1986; Kachel et al., 2002). The resulting ventilation can prevent CO2 from building up in bottom waters (e.g. Cross et al., 2014) and rapidly erode corrosivity. However, stratification induced by river plumes can sometimes overcome and even exacerbate this CO2 accumulation. A fine-scale analysis performed specifically for this synthesis showed that the overlying surface plume was also corrosive (mean ΩAr 0.54 ± 0.15 for S < 25). Freshwater discharge has been shown to have particularly poor buffering capacity (e.g. Cross et al., 2013; Evans et al., 2014), so even diluted surface CO2 concentrations and lowered pCO2 values can result in corrosivity due to diluted TA concentrations. Critically, this indicates that the plume was actively favoring influx of atmospheric CO2 into already corrosive waters.

Atmospheric CO2 influxes to surface waters undersaturated with respect to ΩAr only occur across a small range of DIC and TA concentrations in the low-salinity spectrum due to thermodynamic constraints (Supplemental Fig. 2). However, while the plume remains concentrated, this unique chemistry can act to maintain high CO2 concentrations in the underlying bottom layer and prevent CO2 losses to the atmosphere, resulting in efficient transport of respired terrestrial CO2 to at least the mid-shelf regions. This mechanism may be integral to the transport of corrosive waters to the Chukchi Sea. Over mid-shelf regions (i.e. > 40 m), this stratification eases and proximity to nutrient supply at the shelf break favors local primary production at the surface. This creates a secondary, mid-shelf insulation against CO2 outgassing from bottom waters.

The overlap between these two insulating processes—river water discharge and primary production—was observed during both the ISSS and the RUSALCA programs. The influences of mid-shelf primary production are apparent at the northern end of the Lena River section shown in Fig. 4. Bates (2015) found moderate freshwater influences (e.g. mean S = 27.75 ±/- 0.38 for 0–15 m) north and south of Wrangel Island in conjunction with low surface pCO2 values relative to the atmosphere. In this case, low pCO2 values occurred in conjunction with nutrient drawdown and oxygen supersaturation, indicating moderate local primary production. While these surface waters were not simultaneously corrosive, as near the Lena River plume, the low pCO2 values did insulate the underlying corrosive bottom waters during
downstream transit into the Chukchi Sea. As a result, some corrosive inflow resulting from terrestrial organic matter respiration and the unique physical characteristics of river plumes are likely to influence at least the western branch of the Pacific inflow, and could have contributed to the corrosivity on the western side of Herald Canyon observed during the RUSALCA program.

However, this overlap is not ubiquitous. For example, Laptev Strait is extremely shallow and isolated from the mid-shelf by the Lyakhovksy Islands. Here, surface water CO₂ supersaturations were observed throughout the water column, indicating sea-air CO₂ outgassing (Supplemental Fig. 3). Surface water effluxes were also observed in the shallowest waters outside the Kolyma River, where the net influence of river water discharge had begun to weaken. At the edge of the river water plume, Anderson et al. (2010) found that CO₂ gas transfer rates were particularly high. However, just offshore, primary production again creates a surface layer drawdown of CO₂, insulating high CO₂ concentrations in bottom waters from potential outgassing to the atmosphere (Supplemental Fig. 3). At minimum, it is likely that ~60% of this respired terrestrial carbon is sequestered. Net primary production over the East Siberian Sea is estimated at ~13 Tg C yr⁻¹ (Codispoti et al., 2013), based on non-carbon nutrient drawdown. A synthesis analysis of CO₂ fluxes in the region indicates ~8 Tg C yr⁻¹ ingasses from the atmosphere (Laruelle et al., 2014). To balance the annual primary production with the annual CO₂ influx requires a loss of ~5 Tg C yr⁻¹ back to the atmosphere. Assuming that 100% of this balance comes from outgassing of respired terrestrial carbon in order to scale a minimum sequestration value, we estimate ~9 Tg C yr⁻¹ (61.5%) of the total 14 Tg C yr⁻¹ respired terrestrial carbon load estimated by Gustafsson et al. (2017) is retained in shelf waters for downstream transport.

3.3. Formation of corrosive water in the Chukchi Sea: the biological pump

Rates of primary production on the Chukchi Sea shelf are extremely rapid given the high nutrient supply from the Pacific inflow, exceeding 300 g C m⁻² yr⁻¹ (e.g. Hansell et al., 1993; Hill and Cota, 2005), and result in some of the lowest surface water carbon concentrations observed in the open ocean (Bates et al., 2011). Despite high grazing zooplankton biomass (e.g. Reigstad et al., 2008; Questel et al., 2013), the phytoplankton community is dominated by large species that sink quickly (Moran et al., 2005), leading to an uncoupling between primary production and grazing. This process can be especially rapid for ice-associated phytoplankton communities during sea-ice breakup, which favors aggregation and efficient export of the phytoplankton community as it is released from the ice substrate (Tremblay et al., 1989; Riebesell et al., 1991; Buesseler, 1998; Haeckey et al., 1998; Haeckey and Andersson, 1999; Fortier et al., 2002). Given the lack of retention of this carbon as upper trophic biomass, previous work has indicated that up to 75% of total primary production is exported from the mixed layer (Mathis et al., 2007), leading to efficient export of phytoplankton to colder, light-limited bottom waters and sediments where bacterial communities thrive. This carbon is subsequently remineralized, causing a seasonal buildup of CO₂ concentrations and corrosivity in bottom waters, similar to the respiration of terrestrial organic matter discussed above. In the Chukchi Sea, this seasonal corrosivity eventually becomes a characteristic of a new water mass, PWW.

During the ICESCAPE cruise in 2011, this water mass was already starting to form. Previous work has indicated that this area was the leading edge of winter water formation for the 2011 year (Lowry et al., 2015; Pickart et al., 2016), providing an ideal case to study how corrosivity builds up in PWW. Cool (~−1.6 °C) PWW was observed in conjunction with nutrient- and carbon-replete and low oxygen saturation conditions indicating respiration (Fig. 5). However, these conditions did not correlate precisely. Nitrate concentrations near-bottom varied more closely with density than with oxygen saturation or carbon parameters. Of the chemical products of respiration, retention and transport of nitrate is likely most efficient. Carbon and oxygen respiration signals are likely eroded through mixing and sea-air gas exchange, especially during fall storms, and even during ice-covered periods (Cross et al., 2013; Evans et al., 2015; Bates et al., 2014b). As a
result, carbon and oxygen signals correlate more tightly to each other than to nitrate or physical variables in shallow shelf regions. This thermodynamic decoupling of gases from the other winter-modification processes likely accounts for some of the spatial variability observed in these data at the leading edge of PWW formation.

Similar to the breakdown of stratification and sequestration of high CO$_2$ concentrations at the edges of river plumes discussed above, sequestration of gases may not be as efficient at the leading edge of winter water formation during spring. However, away from both the spatial and the temporal edges of this process, the biological pump does more efficiently transport carbon. Strong physical stratification and rapid export occurring at the peak and end of the spring bloom both favor sequestration of respiratory products in the cooling bottom layer. For example, a nutrient study of the ICESCAPE 2010 and 2011 data by Lowry et al. (2015) found that the biological pump was much more active at the interface between winter water and warmer, shallower water with more light availability. This pump continues to operate beyond the June/July spring season and through the summer. During three monthly occupations of the Chukchi Sea shelf, Mathis and Questel (2013) observed a sharp increase in carbon accumulation between August and September, with the increase persisting through October as the bottom layer cooled. The efficiency of carbon pumping into PWW can also vary inter-annually. For example, very limited corrosivity was observed during the 2010 ICESCAPE program, which took place during a similar time of year. However, during 2010, seasonal ice cover persisted for longer over the shelf, both delaying and limiting the formation of winter water (e.g. Gong and Pickart, 2015). During these years, weaker stratification does not provide optimum stability and nutrient supply for phytoplankton primary production. By contrast, during years with extensive open water, vigorous cooling and ice formation result in the formation of extensive amounts of winter water (e.g. Weingartner et al., 2005). Strong stratification at the surface also helps sequester respired CO$_2$ in subsurface layers.

3.4. Shelf-basin exchange of corrosive water

Transport of Pacific-origin water off the northern Chukchi Shelf occurs via several mechanisms and pathways. Each of the main branches of flow from Bering Strait through the Chukchi Sea ultimately drains into the basin (Fig. 1). Much of the water in the western branch forms an eastward-flowing shelfbreak jet along the northern edge of the Chukchi Sea after exiting Herald Canyon (Corlett and Pickart, 2017). However, most summer export occurs through Barrow Canyon, and some Canyon waters are derived from each branch. The Alaskan Coastal Current (ACC) flows through the eastern side of the canyon (Paquette and Bourque, 1974). This water is rapidly transported along the Alaskan coast, and spends the least amount of time over the Chukchi Sea shelf. The central channel branch feeds both sides of the canyon (Pickart et al., 2016). In contrast to the ACC pathway, some of the central channel branch is fed by waters from the slowest transport pathways across the Chukchi Shelf. Some of the western Bering Strait inflow branch is also diverted to the east just north of Herald Shoal (Pickart et al., 2010) and joins the central channel branch before draining into Barrow Canyon, also a very long transport pathway. These slow pathways favor focused deposition, provide ample time for biogeochemical modification during downstream transit, and are therefore likely preconditioned to exhibit corrosive signals.

The water that drains out of Barrow Canyon can take one of several routes. The water on the eastern side of the canyon—dominated by the ACC—tends to follow the isobaths as they bend to the east, forming the Beaufort shelfbreak jet (Nikolopoulos et al., 2009), downstream of the Herald Canyon shelfbreak jet. There is also turbulent outflow from the canyon in the form of eddies and filaments (Pickart and Stossmeister, 2008). Finally, some of the outflow veers to the west forming a current over the Chukchi continental slope. This recently discovered feature has been named the Chukchi Slope Current, which accounts for up to half of the outflowing transport from Barrow Canyon (Corlett and Pickart, 2017). Notably, both the Chukchi and Beaufort shelfbreak jets as well as the Chukchi Slope Current are known to be baroclinically unstable (Pickart et al., 2005; Spall et al., 2008a, 2008b; Corlett and Pickart, 2017), which promotes shelf-basin exchange. Eddies of Pacific-origin are commonly found off the shelfbreak, the most common of which are cold-core anti-cyclones. These features are believed to be a significant source of carbon and nutrient transport from the northern Chukchi Sea (Mathis et al., 2007).

Several sections in the vicinity of Barrow Canyon are shown in Fig. 6. Corrosive conditions associated with PWW are found in each of the transects. The section to the west of Barrow Canyon (upper panel of Fig. 6) shows two extrema of corrosive PWW, one within the eastward-flowing shelfbreak jet closer to shore, influenced by outflow from Herald Canyon and chaotic shelf-basin exchange off the northern Chukchi shelf, and the other within the westward-flowing Chukchi slope current situated farther offshore. The section through the center of Barrow Canyon (center panel of Fig. 6) shows the $\Omega_{AW}$ - undersaturated winter water banked on the western flank of the canyon. Some of this dense water can transpose to the eastern flank and exit the canyon in the Beaufort shelfbreak jet (Pickart et al., 2005), while some of it exits to the west in the Chukchi slope current further offshore, likely influencing the signal shown in the upper panel. The eastern side...
of the canyon contains warmer and fresher Alaska Coastal Water which is largely supersaturated with respect to \(\Omega_{Ar}\). This is likely because the fast-moving current promotes strong mixing which would lead to extremely efficient sea-air CO\(_2\) equilibration and higher \(\Omega_{Ar}\), as discussed earlier.

As the water from both sides of the canyon exits into the Beaufort shelfbreak jet, the surface ACC current insulates the denser water transposed from the western side of the canyon from sea-air exchange. The section to the east of Barrow Canyon (lower panel of Fig. 6) crosses the Beaufort shelfbreak jet. This current carries the corrosive signature of PWW eastward towards the Canadian Arctic Archipelago. Additional biogeochemical conditioning by respiration is limited, as nutrient availability restrains primary production over this narrow shelf. While previous work has considered permafrost degradation as a potential source of labile organic matter to this region, recent work indicates that respiration of permafrost carbon at the shelfbreak is limited (Hilton et al., 2015). Note that this transect extends far into the interior, demonstrating that corrosive PWW ventilates the halocline throughout the Canada Basin.

Another mechanism of shelf-basin exchange along the Beaufort slope is wind-driven upwelling. Easterly winds drive offshore Ekman transport of surface waters, while halocline waters are drawn up onto the shelf (Pickart et al., 2009). Upwelling events are common throughout the year and occur in all ice conditions as long as pack-ice is mobile (Schulze and Pickart, 2012). Recently, new data from the AON-OA program showed that this can lead to rapid outgassing of CO\(_2\) stored in the halocline by earlier shelf-basin exchange (Mathis et al., 2012). Initially, these upwelling events flush the entire shelf with corrosive halocline waters (Fig. 7), but sea-air CO\(_2\) equilibration diminishes this corrosivity over time, partially mitigating the accumulation of carbon and \(\Omega_{Ar}\) undersaturations in the Pacific halocline. Upwelling has also been observed along the Chukchi shelfbreak (Llinás et al., 2009; Spall et al., 2014), though the frequency and commonality of these events are unknown.

### 3.5. Downstream transport of corrosive Pacific waters

The balance between the carbon pumped into the halocline of the Canada Basin from the East Siberian, Chukchi, and Beaufort Sea shelves and the upwelling occurring along these shelfbreaks ultimately determines the net amount of corrosive water fluxed offshore. Using 6 years of mooring data at the mouth of Barrow Canyon, Itob et al. (2012) determined that the annual mean northward transport of PWW was 0.23 Sv, where they defined water as generally colder than \(-1.5\) °C and saltier than 32.0 (which accounts for both regular winter water and hypersaline winter water). Based on an average DIC concentration of 2200 μmol kg\(^{-1}\) in \(\Omega_{Ar}\) undersaturated PWW across this density range (25.8–27.5 kg m\(^{-3}\)), this results in a delivery of 4.94–5.27 Tg C to the western Arctic Ocean in one year. The new moored observations at the AON-OA time-series site synthesized with the AON-OA shipboard data here provide an opportunity to consider how much of this carbon may be lost to sea-air CO\(_2\) gas exchange during upwelling events after departing the shelf, and provide some insights into how efficiently this system can maintain \(\Omega_{Ar}\) undersaturations over time once PWW exits the shelf.

Fig. 8 shows the calculated \(\Omega_{Ar}\) at the AON-OA time-series site. These data were collected at 127 m depth, in the center of the corrosive Pacific halocline layer (Nikolopoulos et al., 2009; see also Fig. 6). Overall, temperature and \(\Omega_{Ar}\) vary closely with each other, with the cold PWW (\(< -0.1\) °C) almost uniformly corrosive, consistent with the water mass diagram shown in Fig. 2. Corrosive conditions were found over the shelf for 290 days, or 80% of the year-long deployment. \(\Omega_{Ar}\) values were below 0.75 (severely corrosive) for 110 days, or 30% of the year-long deployment. Periodic \(\Omega_{Ar}\) supersaturations were found dominantly between October and March of 2012. After April, \(\Omega_{Ar}\) values were consistently undersaturated, and gradually decreased through the following September. \(\Omega_{Ar}\) supersaturations always coincided with the presence of warmer waters and frequently with indicators of upwelling events, even during complete ice cover. This is consistent with the results of Schulze and Pickart (2012) who found that almost 90% of strong easterly wind events, regardless of season and ice cover, resulted in upwelling.

The wind-driven upwelling events in Fig. 8 were identified using along-coast wind speed (from the Pt. Barrow weather station), along-slope velocity, and salinity in the bottom layer (see Section 2.2). A typical event begins when easterly winds exceed 4 m s\(^{-1}\), followed roughly 8 h later by a reversal of the shelfbreak jet. This current usually flows eastward towards the Canadian Arctic Archipelago, but easterly winds readily reverse its direction. Upwelling commences roughly 10 h after the flow reverses, bringing saltier water past the mooring site. The period of upwelling is identified using the GUI when the near-bottom salinity exceeds the monthly mean value (see Schulze and Pickart, 2012 for details). As seen in Fig. 8, the presence of \(\Omega_{Ar}\) supersaturations correlates well with upwelling activity during the winter months (November–March). Most of these events brought Atlantic Water past the mooring site. By contrast, the upwelling events during the warm months did not generally correspond to an increase in saturation. This is because many of those events brought denser PWW past the mooring. This seasonal variation in upwelling (Atlantic Water events in winter, Pacific Water events in summer) is typical of the Beaufort Slope (Lin et al., 2016).

Importantly, during both types of upwelling, PWW is typically advected onto the shelf (Schulze and Pickart, 2012), where it can have substantial impacts on the carbon system. Fig. 7 shows an example of a Pacific Water event that flooded the entire shelf with corrosive waters. Mathis et al. (2012) showed that the 10-day open-water upwelling event captured in Fig. 7 led to substantial CO\(_2\) outgassing (0.18–0.54 Tg C yr\(^{-1}\)), and estimated that four such events could completely balance the net uptake of CO\(_2\) that occurs in the Beaufort Sea during primary production (2–3 Tg C yr\(^{-1}\)). In our dataset, three upwelling events of sufficient duration (> 10 d) occurred during open water periods (July–September; Table 2), alongside four smaller events. If environmental conditions at the surface were similar to those observed by Mathis et al. (2012), we estimate a total CO\(_2\) flux of 0.5–1.7 Tg C across the total 33 days of wind-forced upwelling conditions (Table 2).

Some additional loss of carbon to the atmosphere is likely during the winter months. A number of recent studies have begun to quantify CO\(_2\) fluxes through sea ice. Amid inconsistent results and varying methods for field testing of flux through sea ice, the community has broadly adopted a proxy that linearly reduces gas transfer velocity relative to sea-ice concentration such that: \(\text{Flux(ice)} = \text{Flux} \times (1 - \text{sea-ice concentration})\). This parameterization has recently been confirmed in the Southern Ocean marginal ice zone (Butterworth and Miller, 2016). For ice concentrations at 90–95%, as would be typical of the Beaufort Sea

![Fig. 7. Cross-shelf sections showing \(\Omega_{Ar}\) and pCO\(_2\) (μatm) offshore of the Mackenzie River (inset) during the 2011 AON-OA cruise. Note the low \(\Omega_{Ar}\) and high pCO\(_2\) at the nearshore, indicative of an upwelling event.](image-url)
negative anomalies in zonal easterly 10 m wind speed; negative anomalies in depth-integrated along-isobath zonal current velocity; positive anomalies in depth-averaged salinity across 50–150 m; positive anomalies in $\Omega_{ar}$ and sustained periods of $\Omega_{ar}$ supersaturation. All anomalies are calculated relative to the annual mean value for the variable. Dark colors indicate statistically long periodic upwelling favorable conditions relative to the mean duration. Lighter colors indicate statistically insignificant but above-average event duration. Undetectable and negative anomalies are not shown (blank values).

### Table 2

Number of days of upwelling-favorable conditions indicated by sustained negative anomalies in zonal easterly 10 m wind speed. Numbers are calculated relative to the annual mean value for the variable. Dark colors indicate statistically long periods of upwelling favorable conditions relative to the mean duration. Lighter colors indicate statistically insignificant but above-average event duration. Undetectable and negative anomalies are not shown (blank values).

<table>
<thead>
<tr>
<th>Start Date</th>
<th>Wind Duration (Days)</th>
<th>Velocity Duration (Days)</th>
<th>Salinity Duration (Days)</th>
<th>$\Omega_{ar}$</th>
<th>$\Omega_{ar} &gt; 1$ (Days)</th>
</tr>
</thead>
<tbody>
<tr>
<td>2-Nov</td>
<td>4.68</td>
<td>16</td>
<td>9.06</td>
<td>5.88</td>
<td>3.38</td>
</tr>
<tr>
<td>5-Dec</td>
<td>6.25</td>
<td>4.81</td>
<td>6.94</td>
<td>4</td>
<td>3.56</td>
</tr>
<tr>
<td>23-Dec</td>
<td>16.56</td>
<td>16</td>
<td>12.43</td>
<td>9.38</td>
<td>7.38</td>
</tr>
<tr>
<td>11-Jan</td>
<td>5.8</td>
<td>3.3</td>
<td>3.23</td>
<td>2.19</td>
<td>2.19</td>
</tr>
<tr>
<td>18-Jan</td>
<td>2.75</td>
<td>4.44</td>
<td>6.06</td>
<td>6.13</td>
<td>5.63</td>
</tr>
<tr>
<td>8-Feb</td>
<td>4.44</td>
<td>3.94</td>
<td>3.44</td>
<td>2</td>
<td>1.68</td>
</tr>
<tr>
<td>13-Feb</td>
<td>4.75</td>
<td>4.44</td>
<td>4.31</td>
<td>7.81</td>
<td>5.19</td>
</tr>
<tr>
<td>19-Feb</td>
<td>21.75</td>
<td>19.38</td>
<td>19.31</td>
<td>26.81</td>
<td>26.75</td>
</tr>
<tr>
<td>12-Mar</td>
<td>6.13</td>
<td>3.38</td>
<td>5.81</td>
<td>6.63</td>
<td></td>
</tr>
<tr>
<td>29-Mar</td>
<td>10</td>
<td>13.5</td>
<td>12.5</td>
<td>10.19</td>
<td>1.31</td>
</tr>
<tr>
<td>15-Apr</td>
<td>3.81</td>
<td>2.25</td>
<td>2.69</td>
<td></td>
<td></td>
</tr>
<tr>
<td>24-Jun</td>
<td>2.31</td>
<td>2.75</td>
<td>1.44</td>
<td></td>
<td></td>
</tr>
<tr>
<td>6-Jul</td>
<td>10.06</td>
<td>6.69</td>
<td>6.31</td>
<td></td>
<td></td>
</tr>
<tr>
<td>6-Aug</td>
<td>6.75</td>
<td>4.69</td>
<td>6.5</td>
<td></td>
<td></td>
</tr>
<tr>
<td>26-Aug</td>
<td>3.38</td>
<td>2.56</td>
<td>2.69</td>
<td></td>
<td></td>
</tr>
<tr>
<td>1-Sep</td>
<td>11.13</td>
<td>9.5</td>
<td>10.5</td>
<td>0.63</td>
<td>0.44</td>
</tr>
<tr>
<td>15-Sep</td>
<td>10.38</td>
<td>11.31</td>
<td>11.38</td>
<td></td>
<td></td>
</tr>
<tr>
<td>11-Oct</td>
<td>1.69</td>
<td>2</td>
<td>1.94</td>
<td>1.56</td>
<td></td>
</tr>
<tr>
<td><strong>AVERAGE</strong></td>
<td>7.53 ± 5.19</td>
<td>6.76 ± 5.02</td>
<td>7.14 ± 4.71</td>
<td>7.03 ± 6.70</td>
<td>4.14 ± 2.02</td>
</tr>
</tbody>
</table>

during the winter months, fluxes would be reduced to 0.5–1% of their open water value. This would result in an additional efflux of 0.01–0.02 Tg C yr$^{-1}$ to the atmosphere.

This loss of carbon due to upwelling events (maximum 1.72 Tg C yr$^{-1}$) balances 9.45–32.26% of carbon transport from winter water. Averaged over the depth of the boundary current and along the area of the entire Beaufort Sea shelf ($9 \times 10^{12}$ m$^2$), this upwelling signal could reduce DIC concentrations in winter water by 184 µmol kg$^{-1}$ on average, which would be more than sufficient to alleviate $\Omega_{ar}$ undersaturations in upwelled waters in this case. This scales well with the data record showing that corrosive conditions are present in the Beaufort Sea boundary current ~80% of the year. It is likely that this corrosivity is persistent except in the presence of direct, episodic influences.

### 4. Discussion

#### 4.1. Ocean acidification in the Pacific Arctic Region

As a global system, the oceans have absorbed approximately 28% of the $\text{CO}_2$ emissions released into the atmosphere by human activity in the last century. Of this anthropogenic carbon, nearly 40% resides in the Pacific Ocean alone (Sabine et al., 2004). Based on previous global and local studies (Sabine et al., 2004; Cross et al., 2013; Le Quéré et al., 2015), a conservative estimate suggests that the North Pacific and Arctic Ocean contain approximately 65 µmol kg$^{-1}$ anthropogenic $\text{CO}_2$. Accordingly, the estimated pre-industrial condition and the present extent of corrosive waters based on our data synthesis are shown in Fig. 9. Near-bottom values are displayed over the continental shelf, while the Pacific halocline layer is displayed in the basin and across the shelfbreak. For most areas of the continental shelf, especially in the Chukchi and Beaufort seas, $\Omega_{ar}$ undersaturations have appeared directly as a result of anthropogenic $\text{CO}_2$, although some areas close to Wrangel Island. Surprisingly, the pre-industrial environment does reflect corrosive conditions downstream of the Lena River outflow. By our calculations, more than double the amount of $\text{CO}_2$ would need to be removed in order to induce $\Omega_{ar}$ supersaturations in this area. While we have considered our 65 µmol kg$^{-1}$ estimate of anthropogenic $\text{CO}_2$ conservative, it is likely that $\Omega_{ar}$ undersaturations are a natural phenomenon in this area and the local ecosystem is likely adapted to these conditions.

The discrete data from 2008 to 2012 show the emergence of corrosive waters across the area over time. The intensity and extent of natural $\Omega_{ar}$ undersaturations observed near the Lena River outflow worsened substantially into the present day ($\Omega_{ar} < 0.5$), while new, unnatural $\Omega_{ar}$ undersaturations have emerged in all three Arctic shelf seas. These patterns are particularly pronounced in areas known to be hotspots for bacterial respiration, where currents are weak and respiration products can build up more efficiently over time. Notably, areas that are extremely shallow (e.g. Hanna and Herald Shoals, Wrangel and Banks Island coasts) and the fast-moving Alaska Coastal Current remain supersaturated with respect to $\Omega_{ar}$. These areas are particularly susceptible to consistent vertical overturning, which likely favors air-sea exchange of $\text{CO}_2$ and does not permit sub-surface buildup of $\text{CO}_2$ and $\Omega_{ar}$ corrosivity.

One of the most notable aspects of Fig. 9 is the spreading of corrosive conditions into the Pacific layer of the halocline since pre-industrial times. This implies that widespread acidification of the halocline is a recent phenomenon driven by anthropogenic $\text{CO}_2$. An anthropogenic $\text{CO}_2$ absorption rate of 0.86 ± 0.12 µmol kg$^{-1}$ yr$^{-1}$ has been observed in the sub-Arctic North Pacific (Watanabe et al., 2011). By linearly extrapolating this rate backwards from the present condition, we estimate that corrosive conditions first appeared in the Pacific halocline between 1975 and 1985, where DIC concentrations were only 24–32 µmol kg$^{-1}$ lower than at present. This coincides closely with estimates that the Bering Sea shelf waters to the south were completely saturated with 45 µmol kg$^{-1}$ anthropogenic $\text{CO}_2$ in 1983 (Chen, 1993), and with recent observations of the buildup of corrosive halocline waters in the northern Chukchi Sea and adjacent Canada Basin (Qi et al., 2016). Importantly, this indicates that the East Siberian Sea has contributed naturally corrosive water to the Pacific halocline through transpolar drift, eddy formation, and other forms of shelf-basin
exchange over time without inducing widespread $\Omega_{Ar}$ undersaturation across the halocline. This natural phenomenon was not enough to induce widespread corrosivity in the absence of anthropogenic $CO_2$.

Our data indicate that $\Omega_{Ar}$ undersaturations extend all the way into the Canadian Arctic Archipelago, including the entrances to Amundsen Gulf and M’Clure Strait. While this synthesis does not extend further, other researchers have traced this Pacific-origin carbon through the entire length of the Archipelago and into the North Atlantic (Shadwick et al., 2011; Yamamoto-Kawai et al., 2013). Just as expanding $\Omega_{Ar}$ undersaturations have been observed in the Pacific halocline, it is likely that corrosive conditions will also begin to progress through the Canadian Archipelago. Pacific source waters are already at or near $\Omega_{Ar}$ undersaturation in Baffin Bay and the Labrador Sea (Acsa-Scott et al., 2010), and compounding carbon accumulation factors could have important implications for the highly productive ecosystems of the Canadian Arctic and the multi-billion dollar commercial fish and shellfish industries on the Atlantic side.

4.2. Potential ecosystem vulnerabilities and resilience

As indicated at the start of this paper, mapping corrosivity over the PAR as we have here is critical to understanding potential ecosystem impacts. OA is projected to have negative physiological effects on many species (Kroeker et al., 2010), including reduced shell or skeletal strength, growth, and integrity to slower growth rates, competitive disadvantage, and reduced survival, especially for some early life history stages with limited coping and internal regulatory capacities (see broad reviews of variable organisal response to OA by Fabry et al., 2008; Doney et al., 2009; Kroeker et al., 2010; Ferrari et al., 2011; Joint et al., 2011; Andersson and Gledhill, 2013). The risk of negative impacts is dependent on the duration, intensity, and extent of acidified conditions, population and life stage exposure to those conditions, and species-specific vulnerability and resilience to these hazards. Here, we have shown the persistent duration and widespread spatial extent of acidified conditions. The newly acidified waters in the eastern East Siberian Sea, the northern Chukchi Sea, the Beaufort Sea shelf break and the inlets of Amundsen Gulf and M’Clure Strait (Fig. 9) certainly represent a significant change in environmental conditions that constitutes new exposure to acidified conditions.

While continuing research will be required to show the species- and population-specific vulnerability to the present and future acidified conditions, bedrock species in the Alaskan food web have already shown vulnerabilities to OA in the laboratory setting. For example, zooplankton like copepods and euphausiids that support large swaths of upper trophic predators from seabirds to whales have shown poor internal ionoregulatory capacity. In response to OA, zooplankton species have displayed reduced hatch rates, delayed embryonic development, and increased post-larval energy demands that could impact over-wintering survival and their energy quality as prey (euphausiids: generally, McBride et al., 2014; Euphausia superba, Kawaguchi et al., 2013; Euphausia pacifica, Cooper et al., 2016 and Mclaskey et al., 2016; copepods: Acartia erythrae and Acartia tonsa, Cripps et al., 2014; Pseudocalanus acuspes, Thor and Oliva, 2015). Seals, walrus, and marine birds (e.g. Erignathus barbatus, Odobenus rosmarus, Somateria fischeri) may be impacted by the inherent vulnerability of calcifying benthic invertebrates and bivalves in their prey base (generally, Kurihara, 2008; Mytilus edulis, C. gigas, Gazeau et al., 2007, 2013; Mercenaria mercenaria and Argopecten irradians, Talmage and Gobler, 2010; Macoma calcarea and Mytilus galloprovincialis, Vithakari et al., 2016; Macoma calcarea, Astarte montagui, and Astarte borealis, Goethel et al., 2017). Early life stages of important commercial species have also displayed laboratory vulnerability to acidified conditions, like larval Arctic and Atlantic cod (Atlantic cod: Gadus morhua, Frommel et al., 2012, Stiasny et al., 2016, and Dahlke et al., 2016; Arctic cod: Boreogadus saida, Kunz et al., 2016) and juvenile (Long et al., 2013b) and adult (Long et al., 2013a) red king and tanner crab (Parathelphus camtschaticus, Chinoecetes bairdi) that could ultimately reduce recruitment to the fished population (Punt et al., 2014).

As shown in Fig. 9, OA exposure is particularly pronounced for two important macroecological environments in the Chukchi Sea: the seasonal ice zone, and the advective pathways between the Pacific and the Arctic Oceans (Carmack and Wassman, 2006; Bluhm et al., 2014; Wassman et al., 2015; Moore et al., 2018). Given the short linkages that characterize the food web in the PAR, OA impacts on important prey species like zooplankton and bivalves can quickly cause a reduction in food supply for corresponding predators. In acidification hotspots like Hanna Shoal in the northern Chukchi Sea (see Figs. 5 and 9), several species of seabirds (especially crested auklets; Aethia cristatella, Kuletz et al., 2015) and walruses (Odobenus rosmarus, Jay et al., 2012 and Kuletz et al., 2015) are found in very high densities during summer and fall. Additionally, the four core-use habitat areas for bowhead whales north of Bering Strait (Balana mysticetus, Citta et al., 2015 and Økken et al., 2018) are all located in seasonally acidified areas identified here. Adveected zooplankton that join the slow transport pathways through
the Central Channel are likely to experience extended exposure to acidified conditions. Even for those organisms that end up on the shelf break of the Chukchi and Beaufort seas through the non-corrosive transport pathways like the Alaska Coastal Current, diurnal vertical migration (Eurytemora affinis, Acartia bifilosa, Almén et al., 2014; Euphausia pacifica, Cooper et al., 2016) and copepod diapause (Calanus spp., Mayor et al., 2015) expose these organisms to the persistently acidified Pacific halocline layer.

Linking these documented species vulnerabilities to the exposure to severe, sustained, and widespread acidified conditions in the Pacific Arctic suggests that important changes may be in store for the Chukchi ecosystem. However, it is critical to note that at the present time, there is limited evidence that exposure to OA has caused serious population declines or change in the Pacific Arctic ecosystem. In part, this may suggest stronger natural resilience than short-term (e.g. single-generation) laboratory studies indicate. Phenotypic and species diversity can represent adaptive capacity in some cases: when one phenotype or species is selectively impacted by OA, the overall population can recover as a better-adapted phenotype or species gains a competitive advantage. While adaptive capacity can be difficult to study in the laboratory, longer-term laboratory studies show that high-latitude copepod populations have demonstrated adaptive capacity to acidified conditions within just two generations of selection (Pseudocalanus acuspes, Thor and Dupont, 2015; De Wit et al., 2015). Relative to the time of emergence of anthropogenically-mediated corrosivity over the last several decades, this adaptation is far more rapid (~180 d), indicating that the current pace of acidification has not overwhelmed this natural plasticity. Additionally, other forms of ongoing Arctic environmental change may actually alleviate OA stress. Sufficient food supplies have been shown to confer resistance to acidification (e.g. corals, molluscs, crustaceans, and echinoderms, Ramajo et al., 2016). If phytoplankton production increases in the Arctic in response to warming temperatures and ice losses (e.g. Arrigo and van Dijken, 2015; Hill et al., 2018) this may be an emerging form of resilience.

However, these small-scale forms of resilience are finite. Metabolic upregulation from increased food consumption can only be sustained up to a certain point (Ramajo et al., 2016; Ries et al., 2009). Especially in a system where species and phenotypic diversity is already limited, natural selection simply creates an even more limited, and therefore less resilient, system. Other environmental stressors represent an additional challenge to understanding this resilience: at this stage, the impacts of acidification on ecosystems are too small to distinguish from inter-annual variations and other forms of climate change already linked to major ecosystem perturbations. Rather than indicating overall ecosystem resilience to acidification, initial indications of resilience may instead simply mask acidification impacts, with much more dramatic results likely to emerge as the system passes critical thresholds. For example, while red king crab larvae and juveniles are vulnerable to OA, models project that the impacts on recruitment and population density will be limited for the next 10–20 years and could be masked by the effects of natural variation in temperature or climate patterns before severe impacts of OA rapidly emerge (Punt et al., 2016).

Given the uncertainty about the balance between ecosystem-level vulnerability and resilience, it is imperative to continue ocean research and monitoring and Alaska-specific laboratory acidification studies across multiple scales in order to support commercial industries, coastal communities, and subsistence fisheries in the PAR. As anthropogenic climate change continues to alter the Arctic environment, both vulnerability and resilience of Arctic ecosystems will continue to change. Acidification stress will continue to worsen; even under the most optimistic emissions scenarios for the next several decades (Bopp et al., 2013; Gattuso et al., 2015), the Chukchi and Beaufort seas are projected to experience persistent surface layer undersaturations beginning in the next few decades (2050–2060; Mathis et al., 2015). Even where acidification is not the primary stress on the environment (e.g. Mayor et al., 2015), it does increase the risk of compounding synergistic impacts from other stressors, like warming, hypoxia, and sea-ice losses (e.g. Gruber, 2011; Byrne and Przeslawski, 2013; Todgham and Stillman, 2013; Kroeker et al., 2013). Many studies predict that the future Arctic will be much different than the one we see today (Grebe, 2012; Doney et al., 2012; Renaud et al., 2015). Evidence-based decision making in the face of these changes has the potential to build adaptive capacity for human communities (Mathis et al., 2015a; IUCN Circumpolar Council-Alaska (ICCA), 2015; Lam et al., 2016) and prevent potential fishery crashes, like the projected reduction in population and catch of crab in Bristol Bay (Punt et al., 2014). A solid research footing will help make those efforts as successful as possible.

5. Conclusions

The oceans of the PAR have rapidly acidified over the last several decades as a result of the natural vulnerability of high-latitude waters, a variety of natural acidification mechanisms, and intrusion of anthropogenic CO2. In this study, we synthesized four recent programs in order to investigate the formation and transport of these corrosive water masses. In the Laptev and East Siberian Seas, river discharge and respiration of terrestrial organic matter was observed to create strong Ωar undersaturations. Our projections of pre-industrial conditions indicate that these Ωar undersaturations likely emerged without anthropogenic influences, although absorption of anthropogenic CO2 substantially increased the duration, intensity, and extent of this corrosivity. In the Chukchi Sea, present-day Ωar undersaturations were found in the slow-moving waters to the north, where accumulation of respiration productions occurring in conjunction with cooling and salinization produced almost uniformly corrosive PWW that ventilates the Pacific halocline layer. However, in shallower areas and in the Alaska Coastal Current, mixing and consistent sea-air exchange prevented strong carbon accumulation and Ωar undersaturations. In the Beaufort Sea, wind-driven upwelling of corrosive Pacific halocline water can completely flush the narrow shelf, producing an episodic OA event but also facilitating the return of a substantial amount of carbon from the ocean to the atmosphere. However, combined with other mechanisms of carbon accumulation, air-sea exchange of CO2 occurring during upwelling events was not sufficient to mitigate Ωar undersaturations.

The corrosive waters formed over the shelves in the PAR were shown to penetrate deep into the Canada Basin and all the way into the Canadian Arctic Archipelago – and hence likely into the North Atlantic. While biological impacts from this recent acidification remain unclear, they could have detrimental effects on ecosystems already undergoing substantial environmental pressure from other forms of global climate change. Prey and predators dependent on the slower advective transport pathways between the Pacific and Arctic Oceans and the seasonal ice zone along the northern Chukchi and Beaufort Shelves may be particularly at risk, given the duration and severity of exposure to seasonal undersaturations. In order to support the management and sustainability of the fisheries in the PAR, it will be critical to continue to monitor global emissions and the rate of OA in the Arctic.

Acknowledgements

We sincerely thank our colleagues involved with the four synthesis programs included in this research. Interdisciplinary work in the Arctic is providing the foundation for closer international collaborations and unprecedented new opportunities. We owe a debt of gratitude to the outstanding officers and crew of the research vessels that make this work possible. We also thank our colleagues who contributed time and energy towards data collection, analysis, and programming, especially Carolina Nobre, Natalie Monacci, Kristen Shake, and Leif Anderson. We are also grateful for our colleagues in the Arctic research community and the Synthesis of Arctic Research (SOAR) program, whose enthusiasm and expertise benefited the preparation, focus, and connections in this manuscript and special issue. SOAR is funded primarily by...


